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## COMPARING OZONE PRODUCTION EFFICIENCY (OPE) OF CHEMICAL MECHANISMS USING CHEMICAL PROCESS ANALYSIS (CPA)

**Final Report** 

May 2025



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# Comparing Ozone Production Efficiency (OPE) of Chemical Mechanisms Using Chemical Process Analysis (CPA)

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### **EXECUTIVE SUMMARY**

Three-dimensional Chemical Transport Models (CTMs) are one of the tools used by regulatory agencies for air quality planning and emissions control strategies. Chemical mechanisms are critical to CTMs and describe the formation of secondary pollutants, such as ozone (O<sub>3</sub>) and particulate matter <2.5  $\mu$ m (PM<sub>2.5</sub>), through chemical reactions of oxidants, organic compounds, and other precursors such as nitrogen oxides (NO<sub>X</sub>). Mechanism intercomparison, especially using metrics which reduce sensitivity to modeling scenarios, is important for interpreting results and assessing uncertainties. In this study, we investigate Ozone Production Efficiency (OPE) as a comparison metric under conditions where ozone production is limited by the availability of NO<sub>X</sub>. OPE is the net number of O<sub>3</sub> molecules produced per NO<sub>X</sub> molecule lost (Kleinman et al., 2002) and is a key descriptor of NO<sub>X</sub>-limited O<sub>3</sub> formation.

The main objectives of this study were to:

- Compute OPE in box model simulations using four widely used chemical mechanisms and compare differences in OPE and O<sub>3</sub> between mechanisms
- Analyze the response of modeled OPE and O<sub>3</sub> to varying anthropogenic emissions
- Review available OPE measurements and compare model results to appropriate data.

We used the Comprehensive Air Quality Model with Extensions (CAMx), configured as a 2-box model, to compute OPE for three Texas locations, Houston-Galveston-Brazoria (HGB), Dallas-Fort Worth (DFW), and San Antonio (SAN). Each model scenario represents typical high-ozone summertime conditions for the locations and input data were developed from 3D CAMx simulations from the Texas Commission on Environmental Quality's (TCEQ) 2019 modeling platform. The CAMx 2-box model scenarios were run using four chemical mechanisms: Carbon Bond version 6 revision 5 (CB6r5; Yarwood et al., 2020), Carbon Bond version 7 revision 1 (CB7r1; Yarwood et al., 2021), Statewide Air Pollution Research Center 2007 (SAPRC07; Carter, 2010); and Regional Atmospheric Chemistry Mechanism version 2 (RACM2; Goliff et al., 2013). We developed new methods to compute OPE in CAMx using chemical process analysis (OPE-CPA) as the ratio of net O<sub>3</sub> production, Pn(O<sub>3</sub>), to net NO<sub>2</sub> production, Pn(NO<sub>2</sub>), where NO<sub>2</sub> is the sum of NO<sub>x</sub> oxidation products, including nitric acid (HNO<sub>3</sub>), peroxyacyl nitrates (PANs), and organic nitrates (ONs).

Our results show that OPE-CPA can be used successfully to compare mechanisms under  $NO_X$ limited conditions. Comparison of OPE-CPA, along with concentrations of important species such as  $O_3$ , nitrogen species, and oxidants, provide a better understanding of how  $O_3$  formation varies between mechanisms, which can support the validity of air quality planning. The four mechanisms considered in this work show a similar response of both OPE-CPA and  $O_3$  to varying anthropogenic emissions of volatile organic compounds (VOCs) and  $NO_X$ . OPE-CPA generally increases with increasing VOC and decreases with increasing NO<sub>X</sub>. O<sub>3</sub> response surfaces for all mechanisms show that the base emissions simulations for 2019 are in a NO<sub>X</sub>-limited regime and the location of the 'ridgeline', which separates NO<sub>X</sub>-limited from VOC-limited conditions, is in a similar location with all mechanisms. There are differences between the mechanisms, however. RACM2 consistently produces the largest OPE-CPA, O<sub>3</sub>, and OH, which is due partially to a slower OH + NO<sub>2</sub> reaction rate constant. HNO<sub>3</sub> is the dominant NO<sub>Z</sub> species in all mechanisms, highlighting the importance of the OH + NO<sub>2</sub> reaction in NO<sub>Z</sub> net production, and consequently OPE, for all mechanisms. Uncertainty in the OH + NO<sub>2</sub> reaction rate is therefore especially important to resolve due to its strong influence on O<sub>3</sub> and OPE. OPE-CPA differences between mechanisms are also greatest at high VOC/NO<sub>X</sub> ratios, indicating the differences in the treatment of NO<sub>X</sub> recycling between mechanisms may be important under these conditions.

We also reviewed available OPE measurements and provide a comprehensive overview of the methods used to calculate OPE from measurements. We determined that aircraft measurements downwind of Houston which estimate OPE from the slope of a linear regression of O<sub>3</sub> and NO<sub>z</sub> concentrations (OPE-plot) are most comparable to our modeled OPE-CPA. Surface measurements are considered unsuitable for comparison due to uncertainties introduced by deposition, especially of NO<sub>z</sub>. OPE-CPA from the HGB simulations was interpolated from the base model year (2019) to measurement years (2000, 2006, 2013) using emissions trends in Harris County. OPE-CPA is generally consistent with OPE-plot, but direct comparison is difficult due to differences in how deposition impacts the two OPE metrics, and uncertainties in how VOC speciation changed over the time period considered. There are no suitable measurements available near Dallas-Fort Worth or San Antonio. Results from the DFW and SAN simulations are therefore provided for comparison to the HGB simulation but are not compared to measurements.

A significant finding of this work is that OPE-CPA for all mechanisms responds nonlinearly to NO<sub>X</sub> and increases at low NO<sub>X</sub> even though O<sub>3</sub> production decreases. Using OPE to predict ozone response to NO<sub>X</sub> emissions reductions is therefore an oversimplification that will tend to overstate ozone reductions. The nonlinear response of OPE-CPA to NO<sub>X</sub> also conflicts conceptually with OPE-plot, which is derived from a linear relationship over varying NO<sub>X</sub> input. Additional measurement and modeling studies are needed to better understand these findings. Aircraft based OPE measurements downwind of previously studied locations would be useful to test mechanism response to emission reductions, and speciated VOC measurements would help characterize the reactivity of emissions. Applying CPA-OPE in 3D simulations is also feasible and complementary with other methods used to probe 3D model simulations, such as sensitivity analysis. 3D simulations of urban plumes using a fine horizontal grid resolution could investigate why measured OPE is often stable within a plume even when subject to varying NO<sub>X</sub>

emissions. Comparison of OPE-CPA and OPE-plot along pseudo aircraft transects in the same simulated plume would help us better understand if the two provide similar estimates of OPE.

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